

Baseline

Edited by Bruce J. Richardson

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Polychlorinated biphenyls (PCBs) in marine organisms from four harbours in Guam

G.R.W. Denton^{a,*}, Lucrina P. Concepcion^a, H.R. Wood^a, R.J. Morrison^b

^a Water and Environmental Research Institute of the Western Pacific, University of Guam, Mangilao 96913, Guam

^b School of Earth and Environmental Sciences, University of Wollongong, NSW2522, Australia

The marine environment has been a major source of protein for the people of Guam, and despite the influx of imported foods, local inhabitants continue to harvest algae, molluscs, crustaceans, sea cucumbers and many different kinds of fish for sale and for home consumption. The isolation of Guam (located in the tropical western Pacific at 13°28'N, 144°45'E) has largely protected these living marine resources from the adverse effects of global pollution. However, the US military has been prominent on the island since WWII and has contributed significantly to disturbances in environmental quality in and around their bases. Moreover, Guam has undergone extensive urban and commercial development over the last 30 years with particular expansion in the hospitality and tourism industries. Such developments have greatly increased problems associated with waste disposal, pollution and environmental management. As a result, coastal waters around much of the island have suffered some form of degradation in recent years.

The precise impacts of recent human activities in the coastal areas of Guam have received minimal detailed scientific study. While the use and disposal of many chemicals

are known to have occurred, the degree of chemical contamination has not been quantified. Such information is critical if the ecological, recreational and commercial potential of nearshore areas is to be maintained. This paper reports on the PCB status of various marine organisms taken from four harbours located along the western edge of the southern half of the island (Fig. 1a). Clean and contaminated sites were identified from sediment analysis in an earlier investigation (Denton et al., 1997, 2005). Species selected for study were dominant ecosystem representatives and included organisms from various trophic levels (Table 1). Special attention was given to those with bioindicator potential as well as those frequently harvested for human consumption. The vast majority of samples were collected from sites a–g in Apra Harbour, the largest and oldest port in Guam (Fig. 1c). Biota collection sites within each of the other harbours (Agana Boat Basin, Agat Marina and Merizo Pier) are shown in Fig. 1 (inset b, d and e, respectively).

The analytical procedures were adapted from the US EPA SW 846 methods (US EPA, 1996) for the physical and chemical evaluation of solid waste, in addition to those recommended by the NOAA National Status and Trends Program for Marine Environmental Quality (NOAA, 1993). Briefly, tissue homogenates (~3 g) were twice extracted with 20 ml of methylene chloride in 50-ml Teflon

* Corresponding author. Tel.: +671 735 2690; fax: +671 734 8890.
E-mail address: gdenton@uog9.uog.edu (G.R.W. Denton).

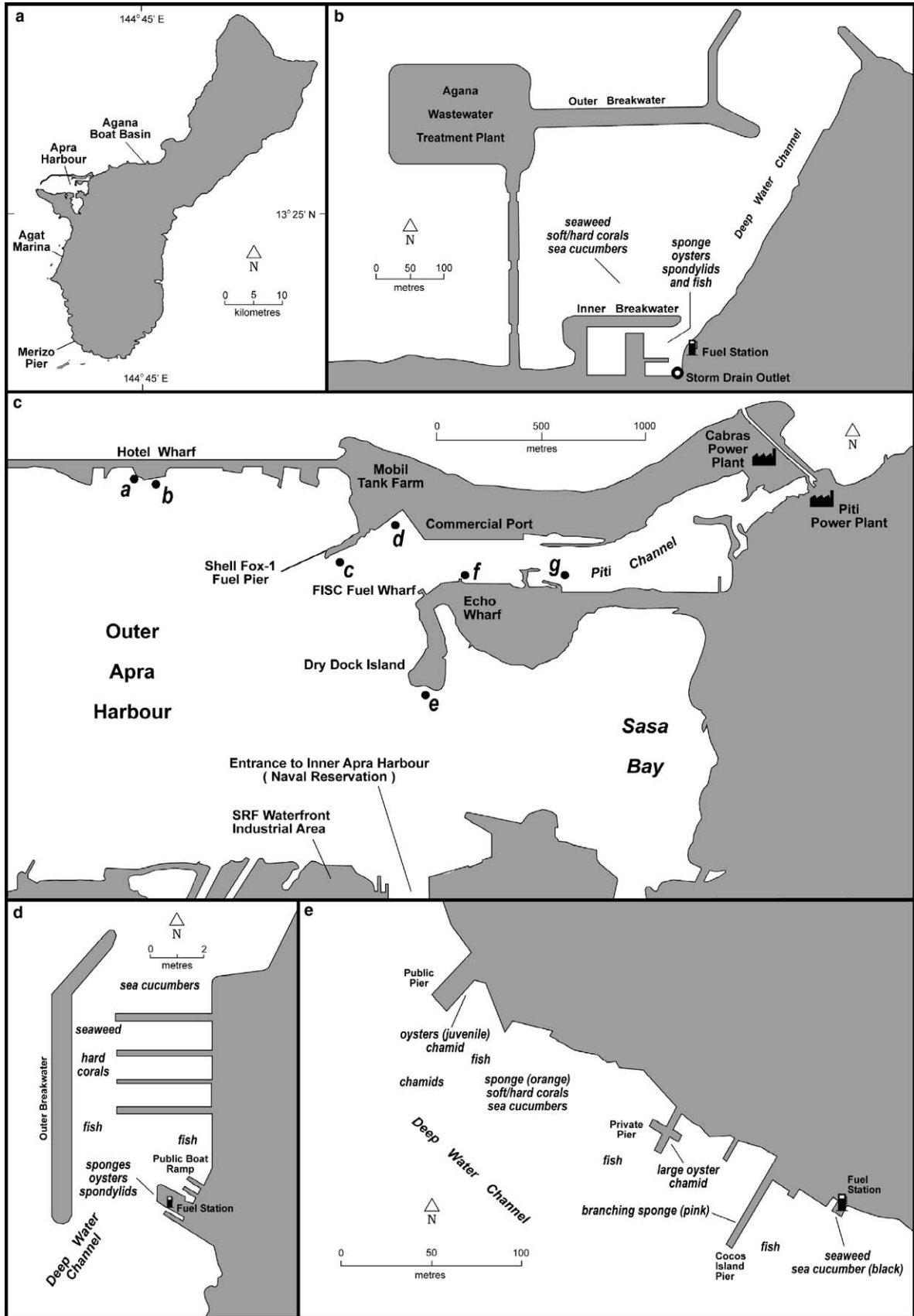


Fig. 1. (a) Locations of harbours studied on Guam and biota sampling sites at: (b) Agana Boat Basin, (c) Outer Apra Harbour, (d) Agat Marina and (e) Merizo Pier.

Table 1 (continued)

Species collected for analysis	Agana Boat Basin	Apra Harbour (site a)	Apra Harbour (site b)	Apra Harbour (site c)	Apra Harbour (site d)	Apra Harbour (site e)	Apra Harbour (site f)	Apra Harbour (site g)	Agat Marina	Merizo Pier
<i>Lutjanus kasmira</i>										x
<i>Monodactylus argenteus</i>	x				x					
<i>Naso annulatus</i>						x				
<i>Naso unicornis</i>		x	x							
<i>Odenus niger</i>									x	
<i>Parupeneus barberinus</i>										x
<i>Parupeneus cyclostomus</i>										x
<i>Parupeneus multifasciatus</i>										x
<i>Saurida gracilis</i>	x								x	
<i>Saurida nebulosa</i>			x							x
<i>Scarus sordidus</i>						x				
<i>Siganus spinus</i>	x									
<i>Sufflamen chrysoptera</i>						x				
<i>Valamugil engeli</i>				x						

Key to Apra Harbour sites: Apra Harbour (site a)—Western end of Hotel Wharf; Apra Harbour (site b)—Central Hotel Wharf; Apra Harbour (site c)—Shell Fox-1 Fuel Pier; Apra Harbour (site d)—Northwestern end of Commercial Port; Apra Harbour (site e)—Southern end of Dry Dock Island; Apra Harbour (site f)—Eastern end of Echo Wharf; Apra Harbour (site g)—Off Port Authority Beach.

centrifuge tubes containing 10 g of anhydrous sodium sulfate and 100 µl of the surrogate congener, PCB 103 (100 pg/µl). After centrifugation, the combined extract was solvent exchanged with hexane and reduced to a final volume of ~0.2 ml under a gentle stream of nitrogen in a water bath (45 °C).

Cleanup was accomplished on 7 mm diameter glass columns of silica gel (2 g) over alumina (1 g) using 5 ml of pentane (discarded) and 10 ml of 50% methylene chloride as the eluting solvents. The extract was solvent exchanged into hexane and reduced to a final volume of 0.1 ml before

adding 10 µl of the internal standard, pentachloronitrobenzene (250 pg/µl). All solvents used were pesticide grade and were checked for interfering contaminants following a 500-fold reduction in volume before analysis.

Analysis was carried out by gas chromatography (Varian 3400CX) using an electron capture detector and a 60 m × 0.25 mm i.d. fused silica MDN-5S, polymethyl-5% phenyl-siloxane (0.25 µm film thickness) capillary column (Supelco). Gas flows (nitrogen) through the column and the detector were 1 ml/min and 30 ml/min respectively. The initial column temperature was maintained at 50 °C

Table 2

Recovery of PCBs from marine mussel standard reference material (SRM 2974) and spiked oyster composite

PCB congener	Certified mean (ng/g) (±95% confidence limits)	This study: (ng/g) mean and (range)	Spike added (ng)	Recovered amount (ng) mean and (range)
<i>SRM 2974: Marine mussel</i>			<i>Oyster composite</i>	
PCB 8	No value	No value	10	11 (9.8–12.1)
PCB 18	26.8 (23.5–30.1) ^a	14.9 (11.6–18.7)	10	9.8 (8.9–10.6)
PCB 28	79 (64–94) ^a	59.2 (41.5–77)	10	13.4 (11.9–15)
PCB 52	115 (103–127)	76.5 (57.1–93.9)	10	5.0 (3.1–6.9)
PCB 44	72.7 (65–80.4)	50.6 (41.1–60.1)	10	12.2 (10.9–13.6)
PCB 66	101.4 (96–106.8)	77.1 (62.1–86.3)	10	12.2 (10.6–13.7)
PCB 101	128 (118–138)	102.9 (75.8–119.1)	10	8.9 (6–11.8)
PCB 77	No value	No value	10	15.8 (13.7–18)
PCB 118	130.8 (125.5–136.1)	125.5 (101.7–144.4)	10	11.1 (9.5–12.7)
PCB 153	145.2 (136.4–154)	92.5 (86.3–103.3)	10	7.5 (6.9–8)
PCB 105	53 (49.2–56.8)	41.6 (36.1–47.6)	10	11.7 (9.9–13.6)
PCB 138	134 (124–144)	65.5 (56.4–77.8)	10	7.2 (6.3–8.3)
PCB 126	No value	No value	10	13.8 (11.2–16.3)
PCB 187	34 (31.5–36.5)	21.1 (17.9–23.3)	10	6.4 (5.1–7.8)
PCB 128	22 (18.5–25.5)	13.1 (10.3–15.1)	10	8.8 (7.5–10.2)
PCB 180	17.1 (13.3–20.9)	7.7 (5.1–9.3)	10	5.5 (4.6–6.5)
PCB 170	5.5 (4.4–6.6)	2.1 (1.2–2.8)	10	6.9 (5.8–8.1)
PCB 195	No value	No value	10	5.6 (4.7–6.5)
PCB 206	No value	No value	10	3.2 (2.5–3.9)
PCB 209	No value	No value	10	1.8 (1.3–2.3)

^a Unconfirmed reference value only.

<i>Haliotis maculata</i>	Bodywall	BDL	BDL	BDL	BDL	0.63	1.89	BDL	0.59	2.07	2.24	0.15	0.22	0.58	BDL	BDL	BDL	10.5
<i>Haliotis maculata</i>	Hemal system	BDL	BDL	BDL	BDL	9.06	591	BDL	4.76	22.9	570	0.29	0.35	3.41	BDL	BDL	BDL	1275
<i>Haliotis maculata</i>	Bodywall	BDL	BDL	BDL	BDL	0.92	0.12	0.73	0.35	2.20	7.00	BDL	0.74	4.75	BDL	BDL	BDL	17.6
<i>Haliotis maculata</i>	Hemal system	BDL	BDL	BDL	BDL	0.97	1.21	0.30	0.12	1.82	3.82	BDL	BDL	1.78	BDL	BDL	BDL	12.8
<i>Haliotis maculata</i>	Bodywall	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.03	BDL	0.05	BDL	BDL	0.17	BDL	BDL	BDL	0.27
<i>Haliotis maculata</i>	Hemal system	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.08	0.10	BDL	BDL	0.06	BDL	BDL	BDL	0.14
<i>Haliotis maculata</i>	Bodywall	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.06	0.10	BDL	BDL	0.22	BDL	BDL	BDL	0.46
BIVALVES (pool size)																		
Oysters																		
<i>Strosireva cucullata</i> (10)	Whole flesh	BDL	BDL	BDL	BDL	0.62	0.38	1.69	BDL	1.74	6.62	BDL	0.22	2.33	BDL	BDL	BDL	14.2
<i>Strosireva cucullata</i> (juv.) (7)	Whole flesh	BDL	BDL	BDL	BDL	BDL	BDL	0.22	BDL	BDL	BDL	BDL	BDL	0.14	BDL	BDL	BDL	1.30
<i>Strosireva mytiloides</i> (4)	Whole flesh	BDL	BDL	BDL	BDL	1.61	BDL	3.38	BDL	2.17	4.19	0.03	BDL	1.73	BDL	BDL	BDL	14.7
<i>Strosireva mytiloides</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	0.37	BDL	1.01	BDL	1.32	4.33	0.01	BDL	1.23	BDL	BDL	BDL	8.54
<i>Strosireva mytiloides</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	0.31	BDL	0.61	BDL	0.16	2.47	BDL	BDL	0.73	BDL	BDL	BDL	4.79
<i>Strosireva mytiloides</i> (1)	Whole flesh	BDL	BDL	BDL	BDL	0.60	BDL	0.67	BDL	0.18	2.82	BDL	BDL	0.77	BDL	BDL	BDL	5.97
<i>Strosireva mytiloides</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	0.21	BDL	0.23	BDL	0.08	1.23	0.04	0.04	0.33	0.05	BDL	BDL	2.60
<i>Strosireva mytiloides</i> (5)	Whole flesh	BDL	BDL	BDL	BDL	0.29	BDL	0.32	BDL	0.08	1.38	0.02	0.02	0.44	BDL	BDL	BDL	3.22
<i>Strosireva mytiloides</i> (1)	Whole flesh	BDL	BDL	BDL	BDL	BDL	BDL	0.33	BDL	0.33	4.09	0.03	0.02	1.06	BDL	BDL	BDL	10.3
<i>Strosireva mytiloides</i> (4)	Whole flesh	BDL	BDL	BDL	BDL	3.42	BDL	2.33	0.34	0.57	13.2	0.08	BDL	8.21	BDL	BDL	BDL	34.5
<i>Strosireva mytiloides</i> (6)	Whole flesh	BDL	BDL	BDL	BDL	3.12	0.26	2.19	0.30	0.60	5.46	0.07	BDL	5.93	BDL	BDL	BDL	38.8
<i>Strosireva mytiloides</i> (3)	Whole flesh	BDL	BDL	BDL	BDL	4.81	0.62	3.41	0.47	1.07	7.24	0.11	BDL	7.41	0.04	BDL	BDL	47.0
<i>Strosireva mytiloides</i> (1)	Whole flesh	BDL	BDL	BDL	BDL	BDL	BDL	0.69	0.73	BDL	3.51	0.06	BDL	1.09	BDL	BDL	BDL	8.15
<i>Strosireva mytiloides</i> (1)	Whole flesh	BDL	BDL	BDL	BDL	0.47	BDL	1.40	BDL	1.75	6.56	0.08	BDL	2.11	0.03	BDL	BDL	14.5
<i>Strosireva mytiloides</i> (1)	Whole flesh	BDL	BDL	BDL	BDL	BDL	BDL	0.16	BDL	0.12	1.09	0.07	0.04	0.42	0.04	BDL	BDL	2.23
<i>Strosireva mytiloides</i> (1)	Whole flesh	BDL	BDL	BDL	BDL	0.42	0.25	1.31	BDL	1.10	5.23	0.06	BDL	1.65	BDL	BDL	BDL	11.5
<i>Strosireva mytiloides</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	BDL	BDL	0.08	BDL	BDL	0.32	0.02	0.05	0.73	BDL	BDL	BDL	1.20
Channids																		
<i>Chama brevicauda</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	0.37	0.68	BDL	0.10	0.33	0.62	0.09	0.16	0.21	BDL	BDL	BDL	3.36
<i>Chama brevicauda</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	BDL	BDL	0.07	BDL	0.12	0.55	0.07	0.19	0.22	BDL	BDL	BDL	1.21
<i>Chama lazarus</i> (3)	Whole flesh	BDL	BDL	BDL	BDL	BDL	BDL	0.07	BDL	0.06	0.25	0.05	0.12	0.11	BDL	BDL	BDL	0.66
<i>Chama lazarus</i> (3)	Whole flesh	BDL	BDL	BDL	BDL	0.11	BDL	0.17	BDL	0.14	0.30	0.14	0.16	0.12	0.08	BDL	BDL	1.87
<i>Chama lazarus</i> (1)	Whole flesh	BDL	BDL	BDL	BDL	0.09	BDL	0.09	BDL	0.09	0.25	0.04	0.09	0.10	BDL	BDL	BDL	0.95
<i>Chama lazarus</i> (1)	Whole flesh	BDL	BDL	BDL	BDL	0.07	BDL	0.08	BDL	0.06	0.17	0.03	0.06	0.07	BDL	BDL	BDL	0.82
<i>Chama brevicauda</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	0.05	BDL	0.18	BDL	0.10	0.22	0.03	0.06	0.07	BDL	BDL	BDL	0.88
<i>Chama brevicauda</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	0.22	0.38	BDL	0.07	0.13	0.27	0.07	0.10	0.10	0.05	BDL	BDL	1.78
<i>Chama brevicauda</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	0.17	BDL	0.09	BDL	0.18	0.87	0.18	0.42	0.36	0.03	BDL	BDL	2.36
<i>Chama brevicauda</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	0.56	BDL	0.42	BDL	0.21	2.58	0.65	1.27	0.97	0.10	BDL	BDL	7.98
<i>Chama brevicauda</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	0.12	BDL	0.08	BDL	0.09	0.43	0.07	0.17	0.15	BDL	BDL	BDL	1.32
<i>Chama brevicauda</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	0.16	BDL	0.09	BDL	0.05	0.53	0.07	0.17	0.16	BDL	BDL	BDL	1.50
<i>Chama lazarus</i> (1)	Whole flesh	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.03	0.07	0.19	BDL	BDL	BDL	0.29
Spongielike																		
<i>Spondylus? Multinurcaus</i> (2)	Whole flesh	BDL	BDL	BDL	BDL	0.29	BDL	2.00	BDL	1.13	3.09	0.23	0.38	1.03	0.02	BDL	BDL	11.3
<i>Spondylus? Multinurcaus</i> (1)	Whole flesh	BDL	BDL	BDL	BDL	0.83	BDL	1.98	BDL	2.01	10.6	1.15	3.24	4.33	0.08	BDL	BDL	30.5
<i>Spondylus? Multinurcaus</i> (1)	Whole flesh	BDL	BDL	BDL	BDL	2.58	BDL	2.23	BDL	3.45	18.8	1.56	5.01	5.92	0.11	BDL	BDL	44.2
<i>Spondylus? Multinurcaus</i> (4)	Whole flesh	BDL	BDL	BDL	BDL	0.25	BDL	0.19	BDL	BDL	0.50	BDL	0.07	0.36	BDL	BDL	BDL	4.19
OCTOPUS																		
<i>Octopus cyanea</i>	Tentacle	BDL	BDL	BDL	BDL	0.23	0.34	1.01	0.12	0.22	2.49	0.56	1.34	0.98	0.07	BDL	BDL	8.78
	Liver	BDL	BDL	BDL	BDL	1.58	8.09	28.8	0.63	3.97	742	BDL	414	22.0	BDL	BDL	BDL	1271
<i>Mantis shrimp</i>	Tail muscle	BDL	BDL	BDL	BDL	0.20	BDL	1.42	0.22	0.43	17.0	2.18	7.37	4.96	BDL	BDL	BDL	38.2
<i>Concodylus</i> sp.																		

(continued on next page)

<i>Epiplatys insidiator</i> (24.5)	Apra Harbour (c)	3-Jun-98	Axial muscle	0.29	BDL	0.07	BDL	0.59	BDL	BDL	0.62	1.65	6.26	0.14	0.15	4.39	8.24	0.80	2.12	1.69	0.09	BDL	BDL	27.1
<i>Epiplatys insidiator</i> (16.0)	Apra Harbour (e)	12-Jun-98	Axial muscle	0.33	BDL	0.19	0.17	BDL	BDL	BDL	0.35	0.25	1.29	BDL	0.14	3.25	15.2	2.51	7.61	4.94	0.39	BDL	0.22	36.8
<i>Epiplatys merra</i> (24.0)	Merizo Pier	22-Dec-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.03	0.07	BDL	0.03	0.09	0.13	0.06	0.14	0.05	BDL	BDL	BDL	0.59	
<i>Gerres argyreus</i> (24.0)	Agana Boat Basin	30-Dec-98	Axial muscle	0.41	BDL	1.43	0.30	1.13	BDL	BDL	0.72	BDL	0.84	BDL	0.16	0.86	1.72	0.23	0.74	0.60	BDL	BDL	BDL	9.13
			Liver	2.19	BDL	78.8	10.1	47.9	BDL	BDL	30.7	13.4	39.8	BDL	4.52	31.2	333	BDL	20.3	18.5	0.56	0.53	BDL	632
<i>Gerres argyreus</i> (15.5)	Agana Boat Basin	30-Dec-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.25	BDL	BDL	0.33	0.50	0.17	0.28	0.15	BDL	BDL	BDL	1.67
<i>Gerres argyreus</i> (16.5)	Apra Harbour (d)	9-Jun-98	Axial muscle	0.37	BDL	BDL	BDL	0.78	BDL	BDL	1.48	1.17	3.26	BDL	0.57	2.24	2.98	0.36	0.78	0.70	BDL	BDL	BDL	14.7
<i>Gerres argyreus</i> (15.0)	Apra Harbour (d)	9-Jun-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.64	0.30	1.04	BDL	0.24	1.15	2.59	0.45	1.26	1.01	BDL	BDL	BDL	8.67
<i>Gerres argyreus</i> (14.5)	Apra Harbour (d)	9-Jun-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.12	0.19	0.40	BDL	0.15	0.57	0.95	0.20	0.54	0.37	BDL	BDL	BDL	3.49
<i>Gymnathorax javanicus</i> (60.0)	Apra Harbour (a)	5-Jun-98	Axial muscle	0.19	BDL	BDL	BDL	0.60	BDL	BDL	0.42	BDL	0.61	0.05	0.19	0.54	2.07	0.25	0.70	0.68	0.02	BDL	BDL	6.33
			Liver	BDL	BDL	BDL	BDL	1.18	BDL	BDL	2.00	BDL	3.01	BDL	BDL	2.61	10.2	1.38	3.13	3.24	BDL	BDL	0.18	27.0
<i>Leiognathus equulus</i> (14.0)	Agat Marina	22-Jan-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.15	0.41	0.28	0.28	2.14	6.57	1.52	4.08	2.53	0.20	BDL	BDL	BDL	18.2
<i>Lethrinus rubripercaulatus</i> (24.5)	Agat Marina	21-Dec-98	Axial muscle	BDL	BDL	0.14	BDL	BDL	BDL	BDL	0.10	0.17	0.51	BDL	BDL	0.77	1.38	0.35	1.09	0.70	0.07	BDL	BDL	5.27
			Liver	BDL	BDL	5.89	BDL	BDL	BDL	BDL	BDL	BDL	14.7	9.55	BDL	32.2	38.8	17.6	17.8	10.8	1.55	BDL	BDL	149
<i>Lethrinus rubripercaulatus</i> (20.5)	Merizo Pier	22-Dec-98	Axial muscle	0.40	BDL	BDL	BDL	BDL	BDL	BDL	0.17	0.11	0.32	BDL	0.07	0.23	0.29	0.04	0.06	0.06	BDL	BDL	BDL	1.74
<i>Lutjanus kasmira</i> (13.5)	Merizo Pier	22-Dec-98	Axial muscle	0.68	BDL	0.11	BDL	0.20	BDL	BDL	BDL	BDL	BDL	BDL	0.10	BDL	0.32	0.05	0.12	0.22	BDL	BDL	BDL	1.81
<i>Monodactylus argenteus</i> (14.5)	Agana Boat Basin	18-Dec-98	Axial muscle	0.41	BDL	1.02	0.53	0.92	1.09	BDL	0.74	0.62	1.28	BDL	0.26	0.56	1.27	0.25	0.68	0.62	0.03	BDL	BDL	10.3
<i>Monodactylus argenteus</i> (17.8)	Apra Harbour (d)	9-Jun-98	Axial muscle	0.60	BDL	0.08	BDL	0.80	BDL	BDL	0.89	0.15	0.73	0.13	0.15	1.10	3.95	0.37	1.13	1.34	0.03	BDL	BDL	11.5
			Liver	0.26	BDL	1.20	2.32	8.58	BDL	BDL	23.0	6.16	32.1	BDL	7.11	1449	4938	BDL	2772	2103	1.70	0.98	0.68	11346
<i>Monodactylus argenteus</i> (17.0)	Apra Harbour (d)	9-Jun-98	Axial muscle	BDL	BDL	0.19	0.38	1.85	BDL	BDL	2.37	0.83	2.54	0.13	0.57	2.06	6.34	0.53	1.52	1.71	0.04	BDL	BDL	21.0
			Liver	0.36	BDL	0.86	1.06	6.05	BDL	0.28	18.4	7.18	31.2	BDL	6.30	43.1	2672	13.1	986	39.8	1.03	0.71	0.51	3827
<i>Monodactylus argenteus</i> (17.0)	Apra Harbour (d)	9-Jun-98	Axial muscle	0.21	BDL	0.19	0.18	1.89	BDL	BDL	2.60	0.39	1.82	0.37	0.50	3.47	14.6	1.36	4.17	4.29	0.13	BDL	BDL	36.1
<i>Monodactylus argenteus</i> (17.0)	Apra Harbour (d)	9-Jun-98	Axial muscle	0.74	BDL	0.13	BDL	1.40	BDL	BDL	1.61	0.32	1.42	0.12	0.28	1.51	5.13	0.40	1.15	1.36	BDL	BDL	BDL	15.6
<i>Monodactylus argenteus</i> (16.8)	Apra Harbour (d)	9-Jun-98	Axial muscle	BDL	BDL	0.55	0.51	2.78	BDL	BDL	2.65	0.79	1.64	0.12	0.44	1.44	4.33	0.39	0.99	1.12	0.04	BDL	BDL	17.8
<i>Monodactylus argenteus</i> (16.5)	Apra Harbour (d)	9-Jun-98	Axial muscle	0.31	BDL	0.61	BDL	2.34	BDL	BDL	2.45	0.61	2.29	0.20	0.53	2.21	7.24	0.81	2.09	2.43	0.07	BDL	BDL	24.2
			Liver	1.10	BDL	3.13	2.49	17.0	BDL	BDL	29.0	8.79	38.0	2.64	8.91	38.1	124	13.8	50.7	47.8	1.50	1.67	1.34	390
<i>Naso annulatus</i> (13.5)	Apra Harbour (c)	12-Jun-98	Axial muscle	BDL	BDL	BDL	BDL	1.03	BDL	BDL	0.22	BDL	BDL	BDL	BDL	2.25	0.19	0.88	1.36	0.07	BDL	BDL	0.07	6.08
<i>Naso unicolor</i> (18.5)	Apra Harbour (a)	5-Jun-98	Axial muscle	1.57	BDL	0.06	BDL	0.25	BDL	BDL	0.21	0.04	0.15	0.06	0.07	0.33	1.42	0.19	0.58	0.55	0.02	BDL	BDL	5.51
<i>Naso unicolor</i> (25.0)	Apra Harbour (a)	5-Jun-98	Axial muscle	0.19	BDL	BDL	BDL	0.38	BDL	BDL	0.20	BDL	0.16	BDL	0.06	0.23	1.02	0.13	0.35	0.35	BDL	BDL	BDL	3.06
<i>Odonis niger</i> (17.0)	Agat Marina	22-Jan-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.11	1.22	0.05	0.14	0.09	BDL	BDL	BDL	0.61
<i>Parupeneus barberinus</i> (26.0)	Merizo Pier	22-Dec-98	Axial muscle	0.30	BDL	0.12	BDL	0.29	BDL	BDL	0.14	0.08	0.25	BDL	0.06	BDL	0.38	0.11	0.35	0.21	BDL	BDL	BDL	2.30
<i>Parupeneus barberinus</i> (16.0)	Merizo Pier	22-Dec-98	Axial muscle	0.21	BDL	BDL	BDL	0.23	BDL	BDL	BDL	0.14	BDL	BDL	0.12	0.16	0.02	0.05	0.03	0.03	BDL	BDL	BDL	0.96
<i>Parupeneus cyclostomus</i> (25.0)	Merizo Pier	22-Dec-98	Axial muscle	0.33	BDL	0.09	BDL	BDL	BDL	BDL	0.17	BDL	0.43	BDL	BDL	0.38	0.95	0.09	0.25	0.20	BDL	BDL	BDL	2.90
<i>Parupeneus multifasciatus</i> (17.5)	Merizo Pier	22-Dec-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.10	BDL	0.55	BDL	0.23	0.40	1.00	0.12	0.36	0.26	BDL	BDL	BDL	3.02
<i>Saurida gracilis</i> (23.0)	Agana Boat Basin	30-Dec-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.20	BDL	BDL	BDL	0.24	0.50	0.14	0.35	0.20	BDL	BDL	BDL	1.64
			Liver	BDL	BDL	12.1	4.67	14.1	3.13	BDL	20.4	BDL	56.6	BDL	10.6	49.5	143	BDL	71.7	33.6	2.98	BDL	0.45	423
<i>Saurida gracilis</i> (19.5)	Agana Boat Basin	30-Dec-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.22	BDL	0.06	0.26	0.51	0.14	0.25	0.17	BDL	BDL	BDL	BDL	1.60
<i>Saurida gracilis</i> (16.5)	Agana Boat Basin	30-Dec-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.16	0.39	BDL	0.12	0.59	1.42	0.30	0.68	0.41	0.03	BDL	BDL	BDL	4.11
<i>Saurida gracilis</i> (15.5)	Agana Boat Basin	30-Dec-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.26	BDL	0.61	BDL	0.19	0.73	1.49	0.28	0.73	0.42	0.03	BDL	BDL	4.75
<i>Saurida gracilis</i> (20.0)	Agat Marina	31-Dec-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.10	0.22	0.08	0.25	0.13	BDL	BDL	BDL	0.78
<i>Saurida gracilis</i> (19.0)	Agat Marina	31-Dec-98	Axial muscle	0.26	BDL	BDL	BDL	0.26	BDL	BDL	BDL	0.11	BDL	BDL	0.11	0.30	0.09	0.29	0.21	BDL	BDL	BDL	BDL	1.63
<i>Saurida gracilis</i> (17.5)	Agat Marina	31-Dec-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.09	BDL	0.06	0.06	BDL	BDL	BDL	BDL	0.21
<i>Saurida nebulosa</i> (21.5)	Apra Harbour (a)	5-Jun-98	Axial muscle	BDL	BDL	BDL	BDL	0.28	BDL	BDL	0.47	0.19	0.90	BDL	0.25	1.69	5.74	1.30	3.39	2.93	0.16	BDL	0.14	17.4
<i>Saurida nebulosa</i> (16.5)	Merizo Pier	22-Dec-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.03	BDL	0.04	0.02	BDL	BDL	BDL	BDL	0.09
<i>Scarus sorchilus</i> (16.0)	Apra Harbour (e)	12-Jun-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.04	0.67	0.46	1.48	0.75	0.10	BDL	BDL	3.50
<i>Scarus sorchilus</i> (15.0)	Apra Harbour (e)	9-Jun-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.43	BDL	0.65	0.30	0.04	BDL	0.08	1.50	
			Liver	3.25	BDL	BDL	3.13	7.52	BDL	BDL	1.70	BDL	BDL	3.74	0.64	BDL	27.3	BDL	116	36.9	8.12	14.6	10.8	234
<i>Scarus sorchilus</i> (14.0)	Apra Harbour (e)	12-Jun-98	Axial muscle	0.20	BDL	0.21	0.20	BDL	0.15	BDL	0.19	0.17	0.18	0.17	0.15	0.14	0.52	0.31	0.74	0.37	0.17	BDL	0.19	3.86
<i>Siganus spinus</i> (15.0)	Agana Boat Basin	18-Dec-98	Axial muscle	0.34	BDL	0.58	0.18	0.82	BDL	BDL	0.60	BDL	0.56	BDL	0.15	0.64	0.84	0.14	0.38	0.31	BDL	BDL	BDL	5.54
<i>Sufflamen chrysopterus</i> (17.0)	Apra Harbour (e)	12-Jun-98	Axial muscle	BDL	BDL	BDL	BDL	BDL	BDL	BDL	0.93	BDL	0.41	0.20	BDL	BDL	16.5	0.54	12.8	5.75	0.06	BDL	0.38	37.6
			Liver	1.11	BDL																			

for the first minute of each run. It was then ramped to 150 °C at 30 °C/min, then to 280 °C at 25 °C/min, where it was held for 20 min to give a total run time of 76 min. Both the injector and detector temperatures were held constant at 280 °C and 310 °C, respectively.

PCB quantification was accomplished using a 20-congener calibration standard representing PCB homologues Cl₂–Cl₁₀ (NOAA, 1993). The congeners were selected on the basis of their potential toxicity, bioaccumulation and/or frequency of occurrence in environmental samples. All calculations were based on peak area comparisons of components sharing identical retention times in both sample and standard. The “total” PCB content of the sample was calculated from the sum of the individual congener data ($\sum_{20}\text{PCB}$). Undetectable congeners were set to zero for this process. Congener recoveries from the certified standard reference material (marine mussel: SRM 2974) and a spiked oyster composite were generally within acceptable limits (Table 2). Method detection limits for individual chlorobiphenyls in the standard mix ranged from 0.02 to 0.15 ng/g.

The congener data for all biotic groups are summarized in Table 3 (all as ng/g wet weight) with no adjustments for recoveries. Co-eluting congeners of possible significance are identified in the legend. Selected PCB data for similar and related species from elsewhere are presented in Table 4 for comparative purposes. In the current work, PCB profiles were dominated by the lower chlorinated homologues (Cl₂–Cl₄) in the brown alga, *Padina* sp., and the mid range homologues (Cl₄–Cl₇) in most other organisms analyzed. An overall analysis of congener prevalence and abundance is presented in Table 5. In both instances, rank orders generally mirrored congener abundance in the technical PCB mixtures, Aroclor 1254 and 1260 (De Voogt et al., 1990). PCB 66, a major component of A1254, was the notable exception here. This congener was detected in <10% of all samples analyzed and had an average abundance of <1%.

PCBs 153 and 187 were the two most commonly encountered congeners, respectively detected in 98% and 97% of all samples analyzed. Overall, they were also the most abundant congeners examined representing 27% and 16% of $\sum_{20}\text{PCB}$ estimates respectively. In contrast, the more toxic planar chlorobiphenyl, PCB 77, a trace constituent of commercial mixtures, was the least prevalent (2.7%) and abundant (0.1%) congener detected. It was detected in only five samples at concentrations ranging from 0.25 to 1.58 ng/g. The highest level was confined to the liver of an octopus taken from Apra Harbour. PCB 126, also a potent planar compound and minor component of commercial mixtures (De Voogt et al., 1990), was an order of magnitude more prevalent than PCB 77 and was detected in all biotic groups, except algae and soft corals. Quantifiable levels of this congener ranged from 0.04 to 9.66 ng/g with highest values occurring in the livers of various fish. It is noted here that PCB 158 co-elutes with PCB 126 under the chromatographic conditions employed

during this study. The former congener is a major component of Aroclor 1260 (Ballschmiter et al., 1987) and one previously reported, albeit infrequently, in environmental samples (McFarland and Clarke, 1989). Consequently, it is possible that the apparent widespread occurrence of PCB 126 in our samples was due in part to the co-presence of PCB 158.

Rank orders of prevalence and abundance for the remaining congeners were generally similar. In summary, PCBs 52, 101, 118, 128, 138, 170 and 180 were detected in >50% of all samples and had average abundances of 3.4–11%. PCBs 8, 28, 44, 105, 195 and 209 were found in <50% and their collective average abundance was <10%. Less commonly encountered in <10% of all samples were PCBs 18 and 206 with a collective average abundance of <0.5%.

Total PCB concentrations ($\sum_{20}\text{PCBs}$) measured during this study varied substantially between and within the biotic groups examined. Concentrations in seaweed and ascidians were consistently <5 ng/g, even from sites previously identified as contaminated. Several sponges yielded values in the same order while others ranged between 10 and 100 ng/g. One encrusting variety (*Dysidea* sp.) had extremely high concentrations that approached 10,000 ng/g at a PCB hot spot, near the Fox-1 Fuel Pier, in Apra Harbour (Fig. 1c, site c). $\sum_{20}\text{PCBs}$ found in an unidentified species of soft coral (*Simularia* sp.) from this location were equally impressive and highlight the bioindicator potential of these organisms for future monitoring purposes. It is noteworthy that both *Dysidea* and *Simularia* are rich in triglycerides, which would explain their high bioaccumulation capacity for PCBs, and presumably other lipophilic xenobiotics. *Dysidea* sp. for example has a lipid content of 20–30%.

PCBs in sea cucumbers were distinctly tissue-dependent, with the hemal system frequently containing levels an order of magnitude or more above those found in body wall muscle. The indigenous people of the western Pacific commonly eat the latter tissue; hence, the typically low PCB levels encountered here (all <20 ng/g) are of interest from a public health standpoint. The unusually high PCB concentration (1279 ng/g) noted in the hemal system of *Holothuria atra*, from a lightly contaminated site at Merizo Pier, is also significant and suggests these organisms may have bioindicator potential.

PCB concentrations in crustaceans and molluscs from uncontaminated waters usually fall within the range 1–10 ng/g (Sericano et al., 1995; Monod et al., 1995; Everarts et al., 1998). Group representatives examined here also generally fell within this range for $\sum_{20}\text{PCBs}$. Exceedences were noted only among those from the contaminated Dry Dock Island site in Apra Harbour (Fig. 1c, site e) and ranged between 30 and 50 ng/g. The only cephalopod mollusc examined during the study was captured beneath the Fox-1 Fuel Pier and, while $\sum_{20}\text{PCBs}$ in muscular tissue was less than 10 ng/g, the concentration in the liver exceeded 1000 ng/g, in line with that noted earlier for

Table 4
PCB concentrations in selected organisms^a from various tropical and subtropical regions of the world

Species	Location	Total PCBs (ng/g wet wt.)	Reference
<i>Algae</i>			
<i>Padina</i> sp. (frond)	Guam harbours	0.39–1.85 (\sum_{20} PCBs)	This study
Various species	Eastern Sicily coastal waters	6–95 ^b	Amico et al. (1979)
Various species	Lagoon of Venice	2–19b	Pavoni et al. (1990)
<i>Dictyota acutiloba</i>			
	Midway Atoll	7.14 (average; \sum_{20} PCBs) ^b	Hope et al. (1998)
<i>Giffordia breviarticulata</i>	Midway Atoll	4.7 (average; \sum_{20} PCBs) ^b	Hope et al. (1998)
<i>Halophila ovalis</i>	Midway Atoll	1.8 (average; \sum_{20} PCBs) ^b	Hope et al. (1998)
<i>Sea cucumbers</i>			
<i>Bohadschia argus</i> (body wall)	Guam harbours	0.03–12.8 (\sum_{20} PCBs)	This study
<i>Bohadschia argus</i> (hemal system)	Guam harbours	0.28–66.5 (\sum_{20} PCBs)	This study
<i>Bohadschia argus</i> (whole)	Midway Atoll	22 (average; \sum_{20} PCBs) ^b	Hope et al. (1998)
<i>Holothuria atra</i> (body wall)	Guam harbours	0.14–17.6 (\sum_{20} PCBs)	This study
<i>Holothuria atra</i> (hemal system)	Guam harbours	0.46–1279 (\sum_{20} PCBs)	This study
<i>Holothuria atra</i> (whole)	Midway Atoll	1.12 (average; \sum_{20} PCBs) ^b	Hope et al. (1998)
<i>Bivalve molluscs (whole flesh)</i>			
Chamids: <i>Chama brassica</i>	Apra Harbour, Guam	1.21–3.26 (\sum_{20} PCBs)	This study
Chamids: <i>Chama lazarus</i>	Guam harbours	0.29–7.98 (\sum_{20} PCBs)	This study
Chamids: <i>Chama iostoma</i>	Midway Atoll	7.6 (average; \sum_{20} PCBs) ^b	Hope et al. (1998)
Oysters: <i>Saccostrea cucullata</i>	Guam harbours	1.3–14.2 (\sum_{20} PCBs)	This study
Oysters: <i>Striostrea</i> cf. <i>mytiloides</i>	Guam harbours	1.2–47.0 (\sum_{20} PCBs)	This study
Oysters: unidentified	Dominican Republic	3.1–8.2 ^b	Sbriz et al. (1998)
Mussels: <i>Perna viridis</i>	Hong Kong waters	39–305 ^b	Phillips (1985, 1986)
Mussels: <i>Mytillus galloprovincialis</i>	Catalan Coast, Mediterranean	2.19–51.1	Porte and Albaiges (1993)
<i>Cephalopod molluscs</i>			
Octopus: <i>Octopus cyanea</i> (tentacle)	Apra Harbour, Guam	8.78 (\sum_{20} PCBs)	This study
Octopus: <i>Octopus cyanea</i> (liver)	Apra Harbour, Guam	1271 (\sum_{20} PCBs)	This study
Octopus: unidentified (whole)	Saint Paul & Amsterdam Islands, Indian Ocean	2.1–5.0 (\sum_6 PCBs) ^b	Monod et al. (1995)
Cuttlefish: <i>Sepia</i> sp. (whole)	East African waters	3.0 (average; \sum_7 PCBs)	Everaarts et al. (1998)
<i>Crustaceans</i>			
Crab: <i>Macropipus tuberculata</i> (whole)	Catalan Coast, Mediterranean	10.2–90.5	Porte and Albaiges (1993)
Crab: <i>Carcinus mediterraneus</i> (whole)	Mediterranean	1448 (maximum)	Fowler (1987)
Mantis shrimp: <i>Gonodactylus</i> sp. (tail muscle)	Apra Harbour, Guam	38.2 (\sum_{20} PCBs)	This study
Shrimp: <i>Parapenaeus longirostris</i> (whole)	Mediterranean	mostly <30	Fowler (1987)
Shrimp: <i>Metapenaeus ensis</i> (whole)	Tonkin Bay, N. Vietnam	0.5–1.1 ^b	Dang et al. (1998)
<i>Fish (muscle)</i>			
31 spp.	Guam harbours	0.09–85.0 (\sum_{20} PCBs)	This study
Several species	Six tropical Asian countries	0.38–110	Kannan et al. (1995)
6 spp.	Hong Kong waters	<0.001–9.6 (\sum_{51} PCBs)	Chan et al. (1999)
5 spp.	Brisbane River, Australia	nd–2100	Shaw and Connell (1980, 1982)
3 spp.	Catalan Coast, Mediterranean	3.10–482	Porte and Albaiges (1993)
Red mullet: <i>Mullus barbatus</i>	Alexandroupolis, Greece	7.9–14.6	Giouranovits-Psyllidou et al. (1994)
<i>Fish (liver)</i>			
15 spp.	Guam harbours	15.8–17.009 (\sum_{20} PCBs)	This study
Parrot fish: unidentified	Madeira, Portugal	44,000 (\sum_6 PCBs) ^c	Ballschmitter et al. (1997)
Bluefin tuna: <i>Thunnus thynnus</i>	Catalan Coast, Mediterranean	112–275	Porte and Albaiges (1993)

nd—not detected.

^a No comparable data found in the literature for sponges, soft corals or ascidians.

^b Original data (ng/g dry weight) converted to ng/g wet weight using average wet:dry weight ratios determined in present study (i.e., 6.3, 8.3, 6.3, 3.8, 4.5 and 4.3 for algae, sea cucumbers, bivalves, octopus, shrimp and fish muscle respectively).

^c Original data reported as ng/g lipid weight converted to ng/g wet weight using a wet weight:lipid weight ratio of 2.5.

sponges (*Dysidea* sp.) and soft corals (*Simularia* sp.) from this site.

Marine fish from uncontaminated waters usually contain PCBs in the low ng/g range in their axial muscle. In the present study, \sum_{20} PCBs ranged from 0.09 to 85 ng/g

in the 74 specimens analyzed. Of these, 13 fish from Apra Harbour contained concentrations greater than 20 ng/g, while a further 13 fish, predominantly from Apra Harbour and Agana Boat Basin, gave values between 10 and 20 ng/g. About 40 specimens contained between 1 and 10 ng/g,

Table 5
PCB prevalence and abundance data in biota from Guam harbours

Ranking	Prevalence ^a		Average abundance ^b	
	Congener	Frequency (%)	Congener	% of \sum_{20} PCB
1	153	98	153	27
2	187	97	187	16
3	118	91	180	11
4	180	81	118	9.0
5	101	80	138	8.0
6	138	78	52	5.5
7	170	72	101	5.2
8	128	72	8	4.6
9	52	54	170	3.4
10	105	46	128	2.8
11	8	45	28	1.8
12	195	45	105	1.5
13	28	34	126	0.9
14	126	27	44	0.6
15	209	24	66	0.5
16	44	23	195	0.5
17	206	8.6	209	0.4
18	18	7.5	18	0.2
19	66	6.4	206	0.1
20	77	2.7	77	0.1

^a Prevalence = frequency of occurrence (%) of each PCB congener for all samples analyzed.

^b Average abundance = [Sum (% of the \sum_{20} PCB represented by each congener in each sample)]/total number of samples.

with representatives from all four harbours. All fish with concentrations of less than 1 ng/g were exclusively from Agat Marina and Merizo Pier. While PCB concentrations in all fish caught were well below the US Food and Drug Administration tolerance of 2 µg/g (ATSDR, 2000), levels in the great majority (83%) were above the more recent US EPA cancer risk (1:100,000) guideline of no more than 1.5 ng/g in fish consumed on an unrestricted basis (US EPA, 2000).

The livers of 20 fish were analyzed during the present investigation. In all cases, \sum_{20} PCBs greatly exceeded those found in axial muscle. No doubt, the higher lipid content of liver tissue greatly enhances its capacity to act as a reservoir for refractory, lipophilic compounds like PCBs. Concentrations exceeding 10,000 ng/g were found in two fish from Apra Harbour. The first fish, *Caranx melampygus*, a relatively large carnivorous species taken from site e, off Dry Dock Island, contained 17,009 ng/g in its liver, while 11,346 ng/g was measured in *Monodactylus argenteus*, a small omnivorous species captured at site d at the western end of Commercial Port. Chromatograms from both fish were not far removed from the commercial PCB mixture, Aroclor 1260. Several workers have explored the potential of fish liver as an indicator tissue for PCBs (Marthinsen et al., 1991; Pereira et al., 1994; Brown et al., 1998). The idea would seem to have merit providing the foraging ranges of species chosen for such purposes are known beforehand.

Overall, the study indicates moderate PCB contamination in biotic components from Apra Harbour. For sessile species, concentrations generally mirrored distribution pat-

terns noted earlier in sediments (Denton et al., 1997, 2005) highlighting small, localized pockets of contamination in the Commercial Port and Dry Dock Island areas. A similar relationship with sediments was noted for biotic representatives collected from the other harbour sites, all of which were relatively clean by world standards. Clearly, Guam is not free of PCBs despite its remote location. Over the years, the improper disposal of waste oil, electrical equipment and other consumables containing PCBs, together with spillages from capacitors and transformers, have had a major impact on coastal waters around parts of the island. For example, a recent study identified a maximum PCB concentration of 80 ppm (as Aroclor 1254) in whole fish caught next to the seawall of a disused military dumpsite at the southern end of the Orote Point Peninsula, approximately 3 km south of Apra Harbour (ATSDR, 2002). Levels in *Padina* spp. (brown algae) samples collected along this stretch of coastline ranged from 0.19 ng/g at the harbour entrance, to 12.22 ng/g at the seawall (Denton, unpublished data). Additional studies are required to extend the database to other coastal areas around the island, particularly those adjacent to military bases, or close to river mouths, storm water discharge points or sewer outfalls.

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References

- Amico, V., Oriente, G., Piattelli, M., Tringali, C., 1979. Concentrations of PCBs, BHCs and DDTs residues in seaweeds of the east coast of Sicily. *Marine Pollution Bulletin* 10, 177–179.
- ATSDR (Agency for Toxic Substances and Disease Registry), 2000. Toxicological profile for polychlorinated biphenyls (update). Department of Health and Human Services, Public Health Service, Agency for Toxic Substances and Disease Registry. 765 pp. plus appendices. Available from: <<http://www.atsdr.cdc.gov/toxprofiles/tp17.html>>.
- ATSDR (Agency for Toxic Substances and Disease Registry), 2002. Health consultation. Exposure pathway evaluation: Consuming fish and seafood from Orote Point (at Spanish Steps) to Facpi Point containing the Orote seafood advisory area and other areas including Gabgab Beach. Orote Landfill/Agat Bay (USN Marianas) Orote, Guam County, Guam. November 26, 2002. US Department of Health and Human Services, Public Health Service, Agency for Toxic Substances and Disease Registry, Division of Health Assessment and Consultation, Atlanta Georgia 30333. 57 pp. plus appendices.
- Ballschmiter, K.H., Froescheis, O., Jarman, W.M., Caillet, G., 1997. Contamination of the deep-sea. *Marine Pollution Bulletin* 34, 288–289.
- Ballschmiter, K., Rappe, C., Buser, H.R., 1989. Chemical properties, analytical methods and environmental levels of PCBs, PCTs, PCNs, and PBBs. In: Kimbrough, R.D., Jensen, A.A. (Eds.), *Halogenated biphenyls, terphenyls, naphthalenes, dibenzodioxins and related products*. Elsevier, pp. 47–69.
- Bright, D.A., Grundy, S.I., Reimer, K.J., 1995. Differential bioaccumulation of non-ortho-substituted and other PCB congeners in coastal Arctic invertebrates and fish. *Environmental Science and Technology* 29, 2504–2512.
- Brown, D.W., McCain, B.B., Horness, B.H., Sloan, C.A., Sloan, K.L., Tilbury, K.T., Pierce, S.M., Burrows, D.G., Chan, S.-L., Landahl, J.T., Krahn, M., 1998. Status, correlations and temporal trends of chemical contaminants in fish and sediment from selected sites on the Pacific coast of the USA. *Marine Pollution Bulletin* 37, 67–85.
- Chan, H.M., Chan, K.M., Dickman, M., 1999. Organochlorines in Hong Kong fish. *Marine Pollution Bulletin* 39, 346–351.
- Dang, D.N., Nguyen, M.A., Nguyen, C.H., Luu, V.D., Carvalho, F.P., Villeneuve, J.-P., Cattini, C., 1998. Organochlorine pesticides and PCBs in the Red River Delta, North Vietnam. *Marine Pollution Bulletin* 36, 742–749.
- Denton, G.R.W., Concepcion, L.P., Wood, H.R., Morrison, R.J., 2005. Polychlorinated biphenyls (PCBs) in sediments from four harbours in Guam. *Marine Pollution Bulletin* 50, 1133–1141.
- Denton, G.R.W., Wood, H.R., Concepcion, L.P., Siegrist, H.G., Elfin, V.S., Narcis, D.K., Pangelinan, G.T., 1997. Analysis of In-Place Contaminants in Marine Sediments from Four Harbor Locations in Guam. A Pilot Study. Technical Report No. 87, Water and Environment Research Institute of the Western Pacific, University of Guam, 120p. Available from: <<http://www.weriguam.org/reports/>>.
- De Voigt, P., Wells, D.E., Reutergardh, L., Brinkman, U.A.Th., 1990. Biological activity, determination and occurrence of planar, mono- and di-ortho PCBs. *International Journal of Environmental Analytical Chemistry* 40, 1–46.
- Everaarts, J.M., Van Weerlee, E.M., Fischerm, C.V., Hillebrand, Th.J., 1998. Polychlorinated biphenyls and cyclic pesticides in sediments and macro-invertebrates from the coastal zone and continental slope of Kenya. *Marine Pollution Bulletin* 36, 492–500.
- Fowler, S., 1987. PCBs and the environment: the Mediterranean marine ecosystem. In: Waid, J.S. (Ed.), *PCBs and the Environment*, vol. III. CRC Press Inc., Boca Raton, FL, pp. 209–239.
- Giouranovits-Psyllidou, R., Georgakopoulos-Gregoriades, E., Vassilopoulos, V., 1994. Monitoring of organochlorine residues in red mullet (*Mullus barbatus*) from Greek waters. *Marine Pollution Bulletin* 28, 121–123.
- Hope, B., Scatolini, S., Titus, E.J., 1998. Bioconcentration of chlorinated biphenyls in biota from the North Pacific Ocean. *Chemosphere* 36, 1247–1261.
- Kannan, K., Tanabe, S., Tatsukawa, R., 1995. Geographical distribution and accumulation features of organochlorine residues in fish in tropical Asia and Oceania. *Environmental Science and Technology* 29, 2673–2683.
- McFarland, V.A., Clarke, J.U., 1989. Environmental occurrence, abundance, and potential toxicity of polychlorinated biphenyl congeners: considerations for a congener-specific analysis. *Environmental Health Perspectives* 81, 225–239.
- Marthinsen, I., Staveland, G., Skaare, J.U., Ugland, K.I., Haugland, A., 1991. Levels of environmental pollutants in male and female flounder (*Platichthys flesus* L.) and cod (*Gadus morhua*) caught during the year 1988 near or in the waterway of Glomma, the largest river in Norway. I. Polychlorinated biphenyls. *Archives of Environmental Contamination and Toxicology* 20, 353–360.
- Monod, J.-L., Arnaud, P.M., Arnoux, A., 1995. PCB congeners in the marine biota of Saint Paul and Amsterdam Islands, Southern Indian Ocean. *Marine Pollution Bulletin* 30, 272–274.
- NOAA, 1993. National Status and Trends Program for Marine Environmental Quality. Sampling and Analytical Methods of the National Status and Trends Program, National Benthic Surveillance and Mussel Watch Projects 1984–1992. Overview and Summary of Methods, vol. I. National Oceanographic and Atmospheric Administration Technical Memorandum NOS ORCA 71. July 1993. 117 pp.
- Pavoni, B., Calvo, C., Sfriso, A., Orto, A.A., 1990. Time trend of PCB concentrations in surface sediments from a hypertrophic, macroalgae populated area of the Lagoon of Venice. *Science of the Total Environment* 91, 13–21.
- Pereira, W.E., Hostettler, F.D., Cashman, J.R., Nishioka, R.S., 1994. Occurrence and distribution of organochlorine compounds in sediment and livers of striped bass (*Morone saxatilis*) from the San Francisco Bay-Delta estuary. *Marine Pollution Bulletin* 28, 434–441.
- Phillips, D.J.H., 1985. Organochlorines and trace metals in green-lipped mussels *Perna viridis* from Hong Kong waters: a test of indicator ability. *Marine Ecology Progress Series* 21, 251–258.
- Phillips, D.J.H., 1986. Organochlorines in green-lipped mussels (*Perna viridis*) from Hong Kong waters. In: *Proceedings, Second International Workshop on the Marine Flora and Fauna of Hong Kong and Southern China*, Hong Kong, 1986.
- Porte, C., Albaiges, J., 1993. Bioaccumulation patterns of hydrocarbons and polychlorinated biphenyls in bivalves, crustaceans, and fishes. *Archives of Environmental Contamination and Toxicology* 26, 273–281.
- Sbriz, L., Aquino, M.R., Alberto de Rodriguez, N.M., Fowler, S.W., Sericano, J.L., 1998. Levels of chlorinated hydrocarbons and trace metals in bivalves and nearshore sediments from the Dominican Republic. *Marine Pollution Bulletin* 36, 971–979.
- Sericano, J.L., Wade, T.L., Jackson, T.J., Brooks, J.M., Tripp, B.W., Farrington, J.W., Mee, L.D., Readmann, J.W., Villeneuve, J.-P., Goldberg, E.D., 1995. Trace organic contamination in the Americas: an overview of the US national status and trends and the International ‘Mussel Watch’ Programmes. *Marine Pollution Bulletin* 31, 214–225.
- Shaw, G.R., Connell, D.W., 1980. Polychlorinated biphenyls in the Brisbane River estuary, Australia. *Marine Pollution Bulletin* 11, 356–358.
- Shaw, G.R., Connell, D.W., 1982. Factors influencing concentrations of polychlorinated biphenyls in organisms from an estuarine ecosystem.

Australian Journal of Marine and Freshwater Research 33, 1057–1070.
US EPA, 1996. SW-846 Test methods for evaluating solid waste physical/chemical methods. US Environmental Protection Agency, Office of Solid Waste and the National Technical Information Service, Springfield, Virginia, CD-ROM.

US EPA, 2000. National guidance. Guidance for assessing chemical contaminant data for use in fish advisories. Risk assessment and fish consumption limits, third ed., Section 4, risk-based consumption limit tables, vol. 2. Available from: <<http://www.epa.gov/ost/fishadvice/volume2/v2ch4.pdf>>.

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Source and distribution of trace metals in the Medway and Swale estuaries, Kent, UK

K.L. Spencer ^{a,*}, C.L. MacLeod ^b, A. Tuckett ^c, S.M. Johnson ^c

^a *Geography Department, Queen Mary, University of London, Mile End Road, London, E1 4NS, UK*

^b *Arcadis Geharty and Miller International, Unit 2 Craven Court, Willie Snaith Road, Newmarket, Suffolk, CB8 7FA, UK*

^c *Department of Earth and Environmental Sciences, University of Greenwich, Chatham Maritime, Kent, ME4 4TB, UK*

Millions of tonnes of fine-grained sediment are dredged annually from Europe's estuaries to maintain navigation routes and permit access to commercial ports and berths. The majority of this material is removed using traditional dredging techniques and disposed of in landfill or at sea. However, the implementation of the EU Landfill Directive has increased the costs associated with the disposal of dredged material in landfill and there is a growing consensus within Europe to manage fine sediment sustainably and examine the beneficial re-use of dredged material. As a result alternative techniques for both sediment dredging and disposal are being investigated throughout Europe.

A number of fine sediment management techniques are currently in operation in the Medway and Swale estuaries. An extensive traditional dredging programme maintains navigation channels and water injection dredging (WID), a relatively novel dredging technique in the UK, is being used to enable continued access to industrial facilities. In addition, English Nature is keen to support a strategy of beneficial sediment re-use within the Medway, using dredged sediment to re-charge or re-create inter-tidal habitats that are currently under threat from sea level rise and associated sediment loss (English Nature, 1999). Water injection dredging involves the injection of jets of water into bottom sediments in situ, decreasing the cohesion between sediment particles and creating a turbulent water-sediment mixture. This mixture acts as a fluid with extremely low viscosity and is able to flow under the influ-

ence of gravity along the sediment water interface. Environmental concerns still surrounding the use of WID are the potential release of sediment-bound contaminants to the over-lying water column and the re-distribution of contaminated sediments over wider areas (Meyer-Nehls et al., 2000; Sullivan, 2000; Ospar Commission, 2004; Spencer et al., 2005).

Preliminary investigations have indicated that the Medway Estuary currently receives low but appreciable inputs of metal contaminants with numerous point and diffuse sources (Spencer, 2002) and a range of historical contaminant inputs are implied in the vertical distributions of heavy metals in salt marsh sediments (Spencer et al., 2003). However, data are yet to be presented providing a detailed picture of surface sediment quality across the estuary and no data are available for the Swale. Where sediment management practices involve the re-distribution of sediment via WID or the beneficial re-use of dredged sediment in habitat re-creation schemes it is vital that those responsible for the management of sediment have access to detailed and accurate information regarding sediment quality. In addition, the EU Water Framework Directive is likely to require the development of sediment environmental quality standards (EQS) enforceable on a pass/fail basis (Crane, 2003). Hence, in this study we have investigated the source, magnitude and distribution of metals within the surface sediments of the inter-tidal zone in the Medway and Swale estuaries and made assessments regarding sediment quality by comparison with published sediment quality guidelines.

Forty five surface sediment samples were collected from the inter-tidal zone during June and July 2000 (Fig. 1). Redox potential and pH were recorded in situ, once the sediment was returned to the laboratory it was stored at –40 °C until required.

* Corresponding author. Tel.: +44 (0) 207 882 7814; fax: +44 (0) 208 981 6276.

E-mail address: k.spencer@qmul.ac.uk (K.L. Spencer).